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neighbourhood of Algiers: use of the sea urchin,
Paracentrotus lividus, as a bioindicator**

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► **To cite this version:**

Dina Soualili, Philippe Dubois, Pol Gosselin, Philippe Pernet, Monique Guillou. Assessment of seawater pollution by heavy metals in the neighbourhood of Algiers: use of the sea urchin, *Paracentrotus lividus*, as a bioindicator. ICES Journal of Marine Science: Journal du Conseil, 2008, 65 (2), pp.132-139. 10.1093/icesjms/fsm183 . hal-00460169

HAL Id: hal-00460169

<https://hal.univ-brest.fr/hal-00460169v1>

Submitted on 26 Feb 2010

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The use of the sea urchin *Paracentrotus lividus* (L.) as bioindicator of the sea water quality of the Algerian coastal environment

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Abstract

This study assessed marine contamination in heavy metals in the vicinity of the Algiers metropolis by combining chemical and toxicological data using the sea urchin *Paracentrotus lividus*. Metals were analyzed in the sediment and in the sea urchin gonads. The sediment toxicity was assessed by bioassays using sea urchin larval development. This study discriminated a site near the Algiers town as highly polluted in Pb. The contamination levels appeared to be toxic for the *P. lividus* larval development. The occurrence of other metals (Fe, Cd, Cu) was poor compared with that reported in the Mediterranean Sea, with exception of Zn that showed high values in female gonads.

Keywords : Algeria, bioassays, gonads, heavy metals, sediment, *Paracentrotus lividus*.

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1 Introduction

2 The Atlantic-Mediterranean sea urchin *Paracentrotus lividus* (Lamarck)
3 (Echinodermata : Echinoidea) is distributed from Ireland to southern Morocco,
4 including throughout the Mediterranean Sea. This opportunistic species has an
5 important ecological role in different ecosystems especially in Mediterranean Sea where
6 its grazing activity can locally control the dynamics of seaweeds (see review Lozano et
7 al., 1995) and seagrasses (see review Tomas et al., 2004). However human activities in
8 coastal zones can alter *P. lividus* morphology, biology and demography (Harmelin et
9 al., 1981; Delmas et Régis, 1984, Pancucci et al., 1992) and over the long-term can
10 modify its ecological role. The sedentary habit of *P. lividus* and its sensitivity to
11 pollutants has led to use this species as biological-biochemical indicator of local
12 pollution (Warnau et al. 1998, Coteur et al. 2001; Bayed et al. 2005). The sensitivity of
13 embryos and larvae of *P. lividus* prompted its use in embryotoxicity tests (Hagström
14 and Lönning, 1973; Pagano, 1986; Warnau et al. 1996). The embryonic and larval
15 development of sea urchins is one of the toxicity assays used in monitoring and risk
16 assessment programmes (see review Beiras et al., 2003).

17 On the Algerian coasts, *P. lividus* is a dominant species of the shallow water ecosystems
18 where it can reach 25 individuals.m⁻² (Semroud, 1993). It colonizes different biotop like
19 *Posidonia oceanica* and *Cymodocea nodosa* meadows, rocky substrate with photophile
20 algae and overgrazed rocky substrate (Semroud, 1993 ; Guettaf, 2000). It lives from the
21 low-water limit to about 10 m depth (Dieuzeide, 1933 ; Semroud, 1993).

22 Around the Bay of Algiers, the populations are exposed to anthropogenic disturbance.
23 More than 35% of the national littoral population live in the Algiers metropolitan area

1 (between Chenoua and Cap Djinet) or more than 4.3 million persons (950 habitants.
2 km⁻²) in a 115 km long coastal zone, This area is highly industrialized, concentrating
3 about 25 % of the Algerian factories with about 1000 factories in 2001 (Larid, 2003;
4 PAC, 2005) including metallurgical, chemical, pharmaceutical, building material and
5 hydrocarbon industries and in a lower proportion mechanical, electric and electronic
6 engineering industries and food and paper factories. A multitude of small polluting
7 activities, official or not, spreads also along the littoral. Two rivers flow in the bay of
8 Algiers, El Hamiz, and principally El Harrach (970 km⁻²; 6 m³ h⁻¹) which drain the
9 domestic and industrial waste waters of the Algiers city. Among the wastewaters which
10 flow in the bay (225 millions m³ y⁻¹) only 8% are treated (PAC, 2005). The marine area
11 under the influence of the Algiers metropolis catches the highest pollution flow of the
12 Algerian coasts (PAC, 2003) with 100 thousand t y⁻¹ of organic chemical compounds,
13 175 thousand s t y⁻¹ of suspended matter, 1 500 t y⁻¹ of nitrogen and 4 thousand s t y⁻¹ of
14 phosphorus (Larid, 2003). In 2004, 46 beaches of the Bay were prohibited from
15 swimming. The sediments of Algiers Bay are highly polluted in heavy metals,
16 hydrocarbons and organic matter. The main metal detected in the coastal water are Hg,
17 Pb, Cu and Zn (PAC, 2003). The biological effects of the pollution on marine
18 ecosystem are important and the recent program on the management of the coastal
19 environment (PAC) pointed out a biodiversity decrease of 14% in the species of major
20 ecological interest (PAC, 2005).

21 The purpose of the present study was to assess the marine pollution in this area by
22 combining toxicological and chemical data using the sea urchin *P. lividus*. Embryo-
23 larval sediment toxicity bioassays and metal contaminant analysis were used. The metal
24 were analyzed in the sediment. The toxicity of their bioavailable fraction was detected
25 by abnormalities in the *P. lividus* development as this species is a valuable indicator of

1 metal contamination and can accumulate metals as a function of the contamination level
2 of the environment (Warnau, 1998). The results of bioassays were compared to the
3 analysis of metal contamination in the sediments and in the gonads of adult sea urchin
4 inhabiting the sediments previously tested. Three sites were chosen in the vicinity of the
5 Bay of Algiers according to potential sources of pollution. The information obtained
6 will allow the ordination of the sampling stations in term of environmental quality.

7

8 Material and Methods

9

10 Sampling sites

11 Two stations, Algiers Beach and Tamentfoust, were located in the Bay of Algiers and
12 another Sidi Fredj, in the North of the Bay of Bou-Ismaïl (east of the Bay of Algiers)
13 (Fig. 1). The site of Algiers Beach should be the most polluted as it is the nearest of the
14 Algiers town. This site is characterized by a habitat mixing rocks with photophile algae
15 and degraded meadows. The site of Tamentfoust, situated in a half closed creek, is
16 characterised by overgrazed rocks with a sparse *Posidonia oceanica* beds. The Bay of
17 Bou-Ismaïl where the site of Sidi Fredj is located has pollution levels much lower than
18 the levels of the Bay of Algiers (PAC, 2005). More distant from the very industrialized
19 area, it encloses a marine medical centre and shows a better health meadow. But
20 Guehioueche and Zelmat (in PAC, 2005) indicated a progressive meadow regression
21 which could be due to the mechanical and organic disturbance caused by the medical
22 centre.

23

24 Collection and preparation of samples

1 Forty sea urchins of 45-65 mm of diameter were sampled by scuba diving in March
2 2002 from each sites between 1 and 5 m depth. This period correspond to the maturity
3 period of the gonads (Semroud and Kada, 1987 ; Guettaf, 2000) before the main
4 spawning which can partly eliminate metal load in the gonad (Warnau et al., 1998).

5 Just after collection, the sea urchins were measured (ambital diameter) in the laboratory
6 and dissected. The sex was determined and the gonads were dried at 60°C for 48h and
7 stored separately in hermetically-sealed polyethylene containers. All the manipulations
8 were performed with stainless steel instruments.

9 Concomitantly to the sea urchin sampling, the upper 2 cm layer of the sediments
10 inhabited by sea urchins was collected by coring (5 cm diameter). The samples of
11 sediments were placed into sealed polyethylene bags, carried to the laboratory on ice
12 and immediately dried at 60°C for 48h until constant weight then stored in hermetically-
13 sealed polyethylene containers at room temperature.

14

15 Metal analysis

16 The concentration of zinc (Zn), lead (Pb), cadmium (Cd), copper (Cu) and iron (Fe)
17 were measuring in the gonads of the urchins and in the sediment according to the
18 method described by Coteur et al. (2003). As the sediment was very heterogeneous with
19 big gravels, only the fraction with grain sizes < 1mm was retained and considered in
20 this study as the 'total fraction'. Metal concentrations were also measured on the
21 fraction with grain sizes < 63 µm. This fraction allows to estimate the organic matter
22 content of the sediment. Pb, Cd and Cu concentrations were determined by graphite-
23 furnace electrothermic atomic absorption spectrometry using a Varian GTA100
24 SpectrAA640Z spectrometer. Concentration of Zn and Fe were measured by flame
25 atomic absorption spectrometry using a GBC 906AA spectrometer. Accuracy of the

1 method was tested using certified reference material (*Mytilus edulis* tissues, BCR nr
2 278R from the Community Bureau of Reference, Commission of the European Union,
3 Brussels, Belgium). Detection limits for Zn, Pb, Cd, Cu and Fe were 0.002, 0.014,
4 0.001, 0.002 and 0.004 µg of metal per mL of digested sample respectively.

5 The gonads of 10 males and 10 females per site were analysed. For the sediment, 3 to 6
6 replicates of total fraction and 0 to 3 replicates of the fraction < 63 µm were used.

7

8 Embryo-larval bioassays

9 The toxicity of the sediments from the previous three sites was assessed by embryo-
10 larval bioassays in May 2003. Natural sea water used for these tests was collected in
11 front of the marine station of Luc-sur-Mer (Normandy, France), the reference site of the
12 Laboratory of Marine Biology of the U.L.B. (Free University of Brussels, Belgium)
13 where the metal analysis were performed.

14 Sea-water was previously allowed to decant for 48 h, filtered by a 22 µm membrane and
15 maintained at $20 \pm 1^\circ\text{C}$ (FSW, filtered seawater). Adult *Paracentrotus lividus* genitors
16 were collected intertidally from a reference population inhabiting the rocky basins of
17 Morgat (Bay of Douarnenez, Brittany, France). Individuals were transferred in the
18 cultivation system of the marine laboratory of the U.M.H. (University of Mons-Hainaut,
19 Belgium). For the embryotoxicity test, reference sediment was sampled at Wimereux
20 (Nord-Pas-de-Calais, France).

21 The embryo-larval bioassays were performed at the laboratory of Mons-Hainaut. Those
22 bioassays were based on the method described by Coteur *et al.* (2003). Spawning was
23 induced by injection of 20 µl per gram of KCL 0.5N through the peristomial membrane.
24 Gametes from 3 females and 3 males were collected in FSW and their quality (*i.e.*,
25 general eggs' shape and sperm's motility) was checked under Olympus TO41 inverted

1 light microscope. Eggs from each female were fertilised by the pooled sperm from the 3
2 males. Embryos at early gastrula stage (4-5 h after fertilization at $14^{\circ}\text{C} \pm 1^{\circ}\text{C}$) from the
3 3 females were mixed before to be used for the bioassays (this pooling leads to one
4 bioassay per site using the same mixed pool of zygotes). Experiments with embryos
5 were performed in six-well plates (Falcon, ref. 35-3046). One plate was dedicated to
6 each sampled sediments (Alger Plage, Tamenfoust, Sidi-Fredj and Wimereux). Each
7 well of those plates were filled with 0.1 g of dried sediment and 10 ml of FWS was
8 added before the bioassay. One plate filled with FWS was used as control. At the
9 beginning of the bioassay, batches of 250-300 embryos were transferred in each well.
10 After 72 h at $14 \pm 1^{\circ}\text{C}$, larvae were fixed by adding 1 ml of formalin (commercial
11 solution, 35%) in each well and the plates were maintained in an oven at 70°C during 2
12 hours. Plates with fixed larvae were stored at 4°C . The frequency of developmental
13 stages was scored in each well on a random sample of 100 larvae. Developmental stages
14 were scored using an inverted light microscope according to the morphological criteria
15 adapted from Warnau and Pagano (1994). Our larvae were classified in 4 categories:
16 normal plutei (“Normal”, N), retarded plutei presenting a delayed development
17 (“Retarded”, R), abnormal plutei with skeletal malformations and/or gut abnormalities
18 (“Pathologic 1”, P1) and embryos whose development ended at the blastula stage or the
19 gastrula stage (“Pathologic 2”, P2). On our figures, the rates of “Viable” larvae were
20 obtained by summing the rates of “Normal” and “Retarded” larvae from each batch.

21

22 Statistical analyses

23 After arcsin transformation, the developmental rates of normal and viable larvae were
24 compared using a one-way ANOVA followed by a Bonferroni’s test (Zar 1996).

1 Dunnet's tests were used to compare the different rates measured to their corresponding
2 controls (Zar 1996). Significant differences were determined at the 95% level.

3 Contamination levels in gonads were compared by 2-way analysis of variance
4 (ANOVA) and subsequent Tukey HSD multiple mean comparison test (effects: sex and
5 sampling site).

6 The relationship between contamination in gonad and sediment and percentage of viable
7 larvae was studied by factor analysis using the principal-component method with an
8 extraction matrix based on Pearson correlation coefficients and the "varimax" method
9 of factor rotation.

10

11 Results

12

13 Contamination level in the sediment and in the sea urchin gonads

14 The fraction of sediment $< 63 \mu\text{m}$ was more or less abundant according to the site, 0.03,
15 0.26, 1.56 % of the total fraction for Sidi Fredj, Tamenfoust and Algiers Beach
16 respectively. The sediment of Algiers Beach had an obviously higher mud content.

17 Tables 1 and 2 present the heavy-metal concentrations in the sediment. Only the levels
18 in the total sediment have been statistically compared because of the low level of mud
19 in the site of Tamentfoust and Sidi Fredj leading to a too low number of replicates in
20 these both sites. Fe, Pb and Zn concentrations in the total fraction of the sediment
21 differed according to the sampled sites. The site of Algiers Beach was the most
22 contaminated in Pb but the less contaminated in Fe and Zn. Fe concentrations were
23 significantly higher in the sediment from Sidi Fredj.

24 Tables 3 and 4 present the heavy-metal accumulation in the sea urchin gonads. Metal
25 concentrations in the gonads were compared according to sex and sampling site by 2-

1 way ANOVA. Cd and Zn concentrations were significantly higher in female than in
2 male gonads while Cu was more concentrated in male gonads ($p < 10^{-4}$). Pb and Fe
3 concentrations did not significantly differ according to sex. Pb, Cu, and Fe
4 concentrations in the gonads differed according to the sampled sites ($p = 10^{-3}$). Pb and Cu
5 concentrations were significantly higher in sea urchins collected respectively in Algiers
6 Beach and Sidi Fredj (Table 3). Fe concentrations in sea urchins from Tamentfoust
7 significantly differed from those of Algiers Beach. No interaction was significant.

8

9 Embryo-larval development

10 A significant reduction of the percentages of both normal and viable larvae was only
11 observed when embryos were exposed to sediment samples from Alger Beach (Fig. 2 A,
12 B). The low rate of normal larvae induced in response to the sediment from Alger Beach
13 is related to an increase of the rates of abnormal plutei (P1) and retarded plutei (R) (see
14 Table 5). The high rate of retarded plutei induced in response to the sediment from
15 Alger Beach (Table 5) explains why the induced rate of viable plutei remains relatively
16 high (Fig 2B.).

17

18 Integrated data analysis

19 The metal concentrations in sediment, in male gonads, in female gonads and percentage
20 of viable larvae were used in the factor analysis. The variables were distributed along
21 axes representing factors according to the correlation coefficient of the variable with the
22 factors (Fig. 3). The first two factors accounted for 99.9 % of the total variance of the
23 data. On the x-axis, the concentrations in Fe, Cu and Zn in the sediment were opposed
24 to the concentrations in Cd and Pb in the sediment. The relationships between the metal
25 level in the sediment and the metal level in the gonads was dependent on the metal. For

1 Pb the link was strong between sediment and the both gonads, for Cd and Cu it was
2 high between sediment and the male gonads. On the x-axis the percentage of viable
3 larvae was closed to the concentrations in Fe, Cu and Zn in the sediment (e.i located on
4 the other end of the x-axis compared to the Pb and Cd contaminations in sediment and
5 gonads). This suggests a close relationship between Pb and Cd levels in sediment and in
6 gonads and a strongly negative interaction with these latter parameters and the
7 percentage of viable larvae.

8 The y-axis comprised, at the negative end, the levels of Zn in the male and female
9 gonads, and the level of Fe in the female gonads, and at the positive end, the levels of
10 Cd and Cu in the female gonads. The percentage of viable larvae was located on the
11 negative part of this axis suggesting that this parameter was negatively intercorrelated
12 with Cd and Cu levels in the female gonads.

13

14 Discussion

15 The main goal of the present study was to assess the marine pollution in the vicinity of
16 the Bay of Algiers by combining chemical and toxicological data using the sea urchin *P.*
17 *lividus* as bioindicator.

18 No site was distinguishable from others by a generalized higher concentration of metals
19 in the total sediments or in the sea urchin gonads. The sediments of Algiers Beach was
20 separated from the others by the highest concentration in Pb but the lowest
21 concentrations in Zn and Fe, the site of Sidi Fredj by the highest concentration in Fe.
22 Algiers Beach is the most muddy site but its present a significant increase in metal
23 contamination only for a metal, Pb, while usually the load of metal in the sediments
24 increase along with the percentage of the fine fraction. This fact can be due to the low
25 sediment contamination in the other metals (see below). In the gonads, Pb also isolated

1 the population of Algiers Beach from the others. In this site, Fe was low but not
2 different from Sidi Fredj. This latter site presented the highest concentrations in Cu. Cd
3 did not discriminate any site.

4 In the sediment, Cd and Cu values can be considered in the background concentrations
5 of the Mediterranean range. Cd values were ranged between to 0.12 to 0.76 $\mu\text{g g}^{-1}$ dry
6 wt in the total fraction of the sediment while the background concentrations of the
7 Mediterranean were estimated between 0.05 and 1 $\mu\text{g g}^{-1}$ dry wt (EEA, 1999). We note
8 however a higher concentration in the fraction with grain sizes $<63 \mu\text{m}$ of the site of
9 Algiers Beach (1.22 $\mu\text{g g}^{-1}$ dry wt). The Cu values ranged from 4.1 to 6.4 $\mu\text{g g}^{-1}$ dry wt
10 in the total fraction of the Algerian sediment and to 20.9 $\mu\text{g g}^{-1}$ in the fraction with grain
11 sizes $<63 \mu\text{m}$ (Algiers Beach) and the background concentrations of the Mediterranean
12 ranged between 5 to 30 $\mu\text{g g}^{-1}$ (EEA, 1999). The Zn (between 0.02 and 0.08 $\mu\text{g g}^{-1}$ dry
13 wt) and Fe concentrations (between 7.1 and 41.5 $\mu\text{g g}^{-1}$) were very low compared to
14 values usually observed in the Mediterranean Sea ranged from 35 to 150 $\mu\text{g g}^{-1}$ for Zn
15 (Saad et al. 1981, Hoogstraten and Nolting, 1991; Storelli et al. 2001) and $> 10^3 \mu\text{g g}^{-1}$
16 for Fe (Saad et al. 1981, Storelli et al. 2001, Menchi et al., 2002). Only Pb values in the
17 total fraction of the sediment of Algiers Beach (39.6 $\mu\text{g g}^{-1}$) and in the fraction with
18 grain sizes $<63 \mu\text{m}$ of Sidi Fredj (40.6 $\mu\text{g g}^{-1}$) exceeded the background concentrations
19 of the Mediterranean range detected between 5.2 to 23.2 $\mu\text{g g}^{-1}$ (EEA, 1999) or the
20 background (4-17 $\mu\text{g g}^{-1}$) given by the reference tables of NOAA (Buchman, 1999).

21 In the gonads, comparisons with *P. lividus* data available from the literature (Table 6)
22 and especially the study of Warnau et al (1998), confirms the conclusions on the level
23 of metal contamination deduced from the sediment analysis except for Zn. The gonad
24 concentrations in Cd, Cu and Fe from the three sampling locations averaged background
25 concentrations in *P. lividus* gonads. Only the gonad contamination in Zn from the 3

1 sites and in Pb from Algiers Beach were obviously higher than the Zn and Pb
2 background concentrations for this species. The Zn concentrations, not significantly
3 different among the sites, were high, comparable to the concentrations estimated in
4 female gonads of *Sphaerechinus granularis* from the bay of Bay of Brest (between 190
5 and 700 $\mu\text{g g}^{-1}$) (Guillou et al. 2000). This bay is known to be contaminated by this
6 metal which covers the whole roof of the town (Troader, 1995). But in the present study
7 the Zn gonad levels appeared disproportionate with the Zn levels in sediment indicating
8 that sediment possibly is not the most contaminated abiotic compartment. But as Zn was
9 dominant in the gonads of the females, the results about this essential element for
10 animal metabolism must be cautiously interpreted (Hambidge et al., 1986). On the
11 contrary, the high Pb levels in gonads reflected the high Pb levels in the sediment.
12 Among the compared sites, only Rabat (Morocco) presented higher concentrations due
13 to the untreated pottery activity of this town (Bayed et al., 2005).

14 However when metal discriminates the site in terms of contamination, the response given
15 by the sediment can differ from that given by the gonad. The geographical Pb gradient
16 observed in the sediment matches closely the gradient reported by the gonad
17 accumulation, pointing out the site of Algiers Beach as the most contaminated in this
18 metal. But Fe and Zn bioaccumulation did not discriminate the sites in the same way as
19 Fe and Zn sediment concentrations. Three explanations can explain the differential
20 metal pattern: i) the bioavailability of the metals is different ii) essential elements as Fe
21 and Zn have different accumulation patterns than the other metals iii) Fe, Zn and Cu
22 concentrations were too low in the sediment to be suitably expressed in the biological
23 compartment. Although the two first hypotheses are plausible, the third one is clear (see
24 above).

1 The high contamination level in Pb in the site of Algiers Beach expressed by the
2 sediment and gonad analysis was confirmed by a recent study in the water and stream
3 sediment of Oued El Harrach which flows in the Algiers Bay. Levels of 21 to 41 $\mu\text{g g}^{-1}$
4 dry wt, close to the levels detected in the total fraction of the sediment of Algiers Beach
5 ($39.6 \mu\text{g g}^{-1}$) were detected in the sediments in the mouth of the Oued. This pollution
6 was probably caused by the discharge of an Algiers untreated industrial wastewater
7 (Yoshida et al., 2005).

8 The results of the bioassays pointed out Algiers Beach as the only site showing toxicity
9 on sea-urchins embryos with 24.3 % of abnormal larvae (vs 7.9 ± 2.8 % for the 3 other
10 sites) and 47,8 % of retarded larvae (vs 22.56 ± 4.4 % for the 3 other sites). No
11 difference was detected between the two other Algerian sites and the reference site and
12 the control. According to the Kobayashi criteria (1991), the level of normal plutei of
13 Algiers Beach (20.5%) expressed a strong inhibition of the sea urchin development in
14 relation with a high disturbance. The results of these assays appeared negatively related
15 with the Pb concentrations in the male and female gonads which were positively
16 correlated with the Pb concentration in the total fraction of the sediment. The integrated
17 data analysis confirms these conclusions. The high positive relationship between the
18 rate of viable larvae and the Cu and Fe concentrations in the sediment and male gonad
19 and the Zn concentration in the sediment would not express a positive effect of these
20 metals on the larval development but more precisely a lack of negative effect due to the
21 low levels of these metals. Nevertheless the position of the viable larvae suggests a
22 negatively correlation with Cd and Cu levels in the female gonads. Although these
23 contamination levels were low, a possible higher influence of the female gonad on the
24 larval viability would be considered when these gonads were contaminated by metals so
25 toxic as Cd and Cu.

1 The Pb pollution was not spread out as the sediment and the sea urchins of the site of
2 Tamenfoust, a little farther from the Oued El Harrach than the site of Algiers Beach, did
3 not present high Pb accumulation or developmental anomalies. In a semi- closed creek,
4 this former site could be protected from the sea water and sediment flows coming from
5 the Oued. The site of Sidi Fredj considered as less contaminated because more distant
6 from the very industrialized area influenced by Algiers metropolis cannot be
7 distinguished from the site of Tamenfoust according to the criteria used in this study.

8

9 In conclusion, this study confirms the link between developmental abnormalities, metal
10 accumulations in the gonads and metal contamination in the sediment when the metal
11 concentration in sediment is sufficiently high. It accredits the indicator value of *P.*
12 *lividus*. In the present study the analysis discriminates a site and a metal, Pb, but the
13 method could be improved by increasing the number of toxicants analysed and by
14 extending the metal analysis to other sea urchin storage organs as the gut which can
15 better reflect a heavy metal accumulation (Warnau et al, 1998; Guillou et al, 2000).

16

17 Acknowledgments.

18 This paper is a part of the thesis of D. Soualili; this study was carried out in the
19 « Laboratoire de Biologie Marine de l'Université Libre de Bruxelles, ULB », in the
20 « Laboratoire de Biologie Marine de l'Université de Mons-Hainaut, UMH » (Belgium),
21 and ended in the «laboratoire des Sciences de l'Environnement Marin de l'Université
22 de Bretagne Occidentale, UBO » (France). The study in Belgium was funded by a grant
23 of the “Institut des Sciences Agronomiques et Vétérinaires de l'Université de Blida
24 (Algeria)”, the stay in France by a grant of the Algerian Department of Higher
25 Education and Research. D. Soualili gratefully acknowledges Professeur M. Jangoux

1 (ULB) for his hospitality, for access to facilities and helpful discussions. The authors
2 also thank Dr F. Boukroufa for assistance in the field, M. Briand (UBO) for her help in
3 the preparation of figures and M.P. Friocourt (UBO) for fruitful discussions about the
4 writing of the English manuscript.

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Legends of figures

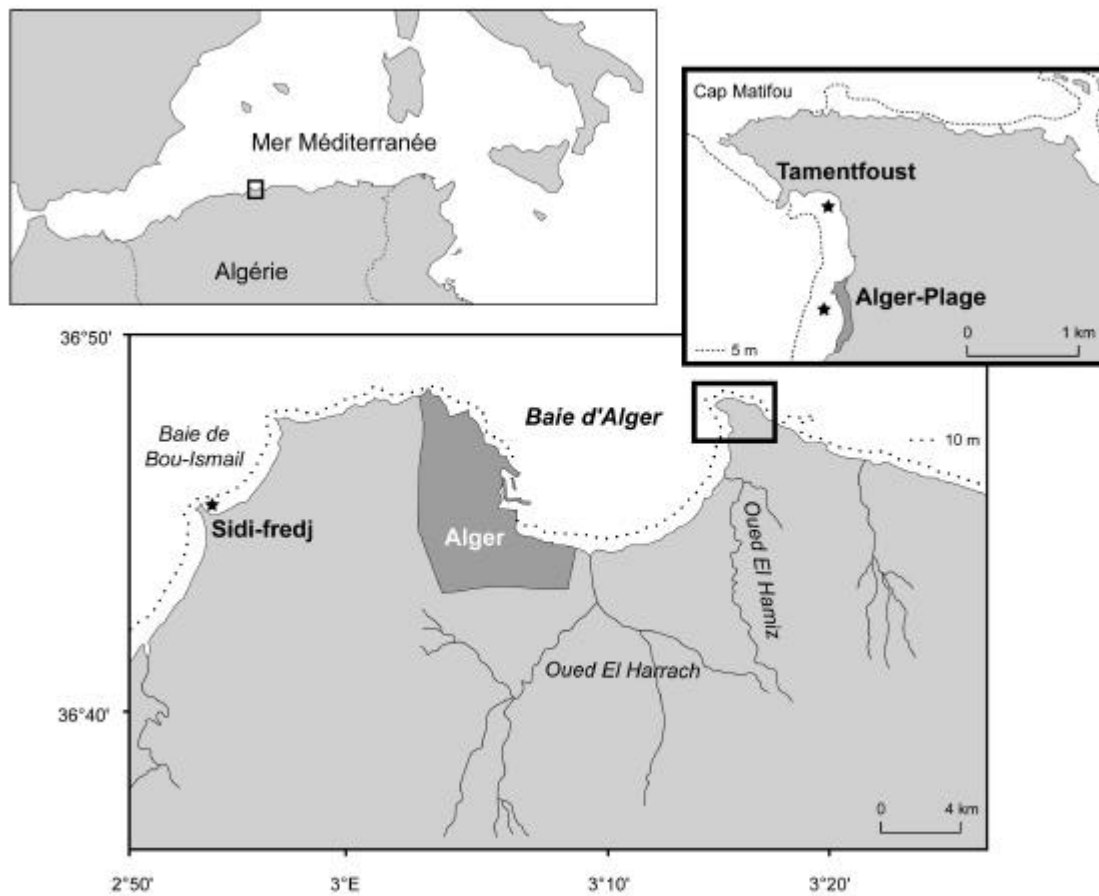


Figure 1 : Location of the three sampling sites: Algiers Beach, Tamenfoust and Sidi Fredj.

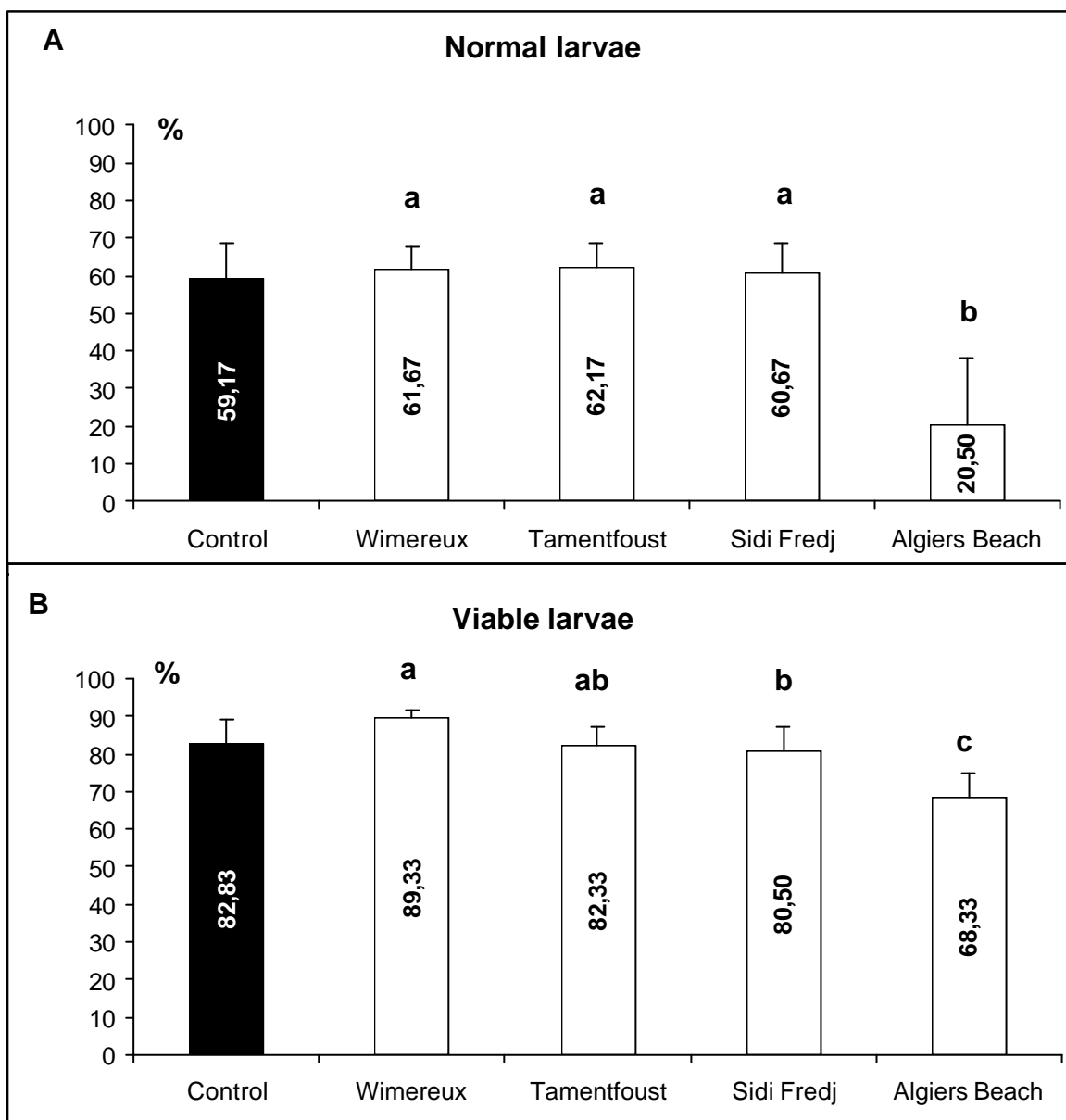


Figure 2. Percentages (mean \pm S.D.) of normal (A) and viable (B) larvae of *Paracentrotus lividus* after exposition to dried sediments and control throughout embryogenesis (72 h). 6 replicates by exposure, 100 larvae scored by replicate. For each species, there were no significant differences between the series designed by the same letter (Bonferroni, $p = 0.05$).

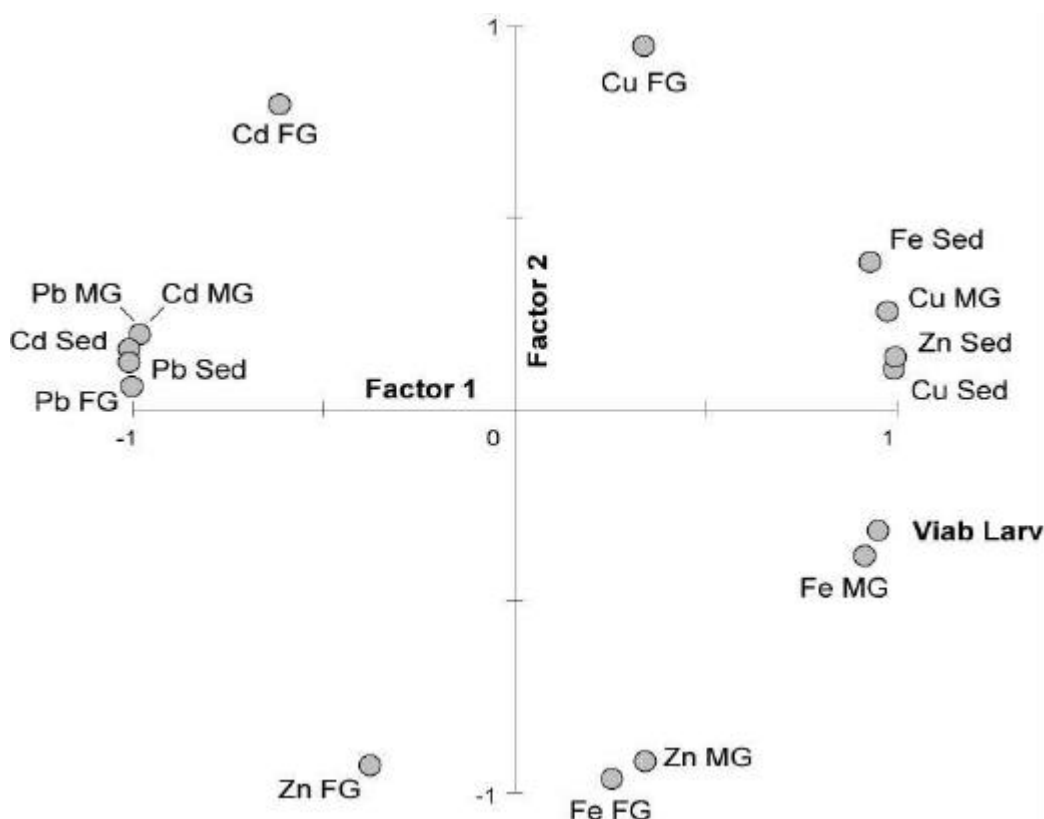


Fig. 3. Factor analysis (principal-component method) showing the relationships between the metal concentrations in the sediment (metal and suffix SED), in the male gonads (metal and suffix MG), in the female gonads (metal and suffix FG) and the percentage of viable larvae (VIABLARV). The first (x-axis) and the second (y-axis) factors accounts 70% and 29.9%. of the total variance respectively.

Legends of tables

Table 1. Metal concentration (mean \pm SD ; $\mu\text{g g}^{-1}$ dw) in the total fraction (n =3 to 6) and in the $<63\mu\text{m}$ grain-size fraction (n= 0 to 3) of the sediment collected in the three Algerian sites (AB: Algiers Beach; TM: Tamenfoust; SF: Sidi Fredj) .

Metals in sediments		Zn	Pb	Cu	Cd	Fe
Total fraction						
AB	Means	0.023	39.63	4.08	0.76	7.11
	SD	0.005	7.93	0.62	1.05	2.18
TM	Means	0.045	14.59	5.73	0.15	17.85
	SD	0.010	1.55	1.65	0.06	3.71
SF	Means	0.050	12.37	6.43	0.12	31.32
	SD	0.006	4.07	1.32	0.01	6.49
<63 μm fraction						
AB	Means	0.081	23.76	20.89	1.22	19.70
	SD	0.008	0.75	0.75	0.04	2.08
SF	Means	0.08	40.63	10.48	0.25	41.50
	SD	-----	-----	-----	-----	-----

Table 2. Comparison of the metal concentrations in the total fraction of the sediments measured in the three Algerian sites (AB: Algiers Beach; TM: Tamenfoust; SF:Sidi Fredj) .

Metal in sediments Total fraction	<i>p</i> ANOVA	Level of contamination ^b		
		+	-	
Zn	0.03	<u>SF</u> <u>TM</u> AB		
Pb	$> 10^{-2}$	AB <u>TM</u> <u>SF</u>		
Cu	NS ^a	TM SF AB		
Cd	NS	<u>AB</u> <u>TM</u> <u>SF</u>		
Fe	$> 10^{-2}$	SF TM AB		

^a NS, no significant difference

^b stations which do not differ in metal concentration are underlined ($p > 0.05$, Tukey HSD test)

Table 3. Metal concentrations (mean \pm SD ; $\mu\text{g g}^{-1}$ dw; n=10) in the gonads of *Paracentrotus lividus* collected in the three Algerian sites (AB: Algiers Beach; TM: Tamenfoust; SF: Sidi Fredj) .

Metals in gonads		Zn	Pb	Cu	Cd	Fe
AB						
Females	Means	385.5*	6.14	2.84	0.14	73.8
	SD	344.1	3.46	0.97	0.08	35.5
Males	Means	32.9	7.78	3.19	0.08	19.3
	SD	13.5	8.77	0.83	0.04	19.7
TM						
Females	Means	538.2*	1.5	2.49*	0.12	113
	SD	324.3	1.72	0.47	0.08	37.6
Males	Means	76.1	0.88	3.88	0.05	112.6
	SD	172.2	0.44	0.84	0.01	66
SF						
Females	Means	366.9	0.68	3.42	0.14*	71.1
	SD	178.3	0.12	0.85	0.09	54.8
Males	Means	52.9	0.90	4.42	0.05	92.7
	SD	73.2	0.41	0.56	0.03	78.8

* indicates a significant difference between the sexes

Table 4 Comparison of metal concentrations in the gonads of *Paracentrotus lividus* measured in the three Algerian sites (2-factor ANOVA: sex and site) (AB: Algiers Beach; TM: Tamenfoust; SF:Sidi Fredj) .

Metal in gonads	Sex	<i>p</i> ANOVA			n	Level of contamination ^b		
		Site	Interaction			+	-	
Zn	<10 ⁻⁴	0.26 ^a	0.53	60	TM SF AB			
Pb	0.69 ^a	<10 ⁻⁴	0.66	60	AB TM SF			
Cu	<10 ⁻⁴	<10 ⁻²	0.11	60	SF TM AB			
Cd	<10 ⁻⁴	0.45 ^a	0.80	60	SF AB TM			
Fe	0.40 ^a	<10 ⁻³	0.06	60	SF AB TM			
						TM SF AB		

^a no significant difference

^b stations which do not differ in metal concentration are underlined ($p > 0.05$, Tukey HSD test)

Table 5. Frequencies (means \pm standard errors) of developmental stages in *Paracentrotus lividus* larvae exposed to sediments samples throughout embryogenesis.

	N	R	P1	P2	Du1	V	Du2
Control	59.2 \pm 9.7	23.7 \pm 5.9	8.7 \pm 4.7	8.5 \pm 3.2		82.8 \pm 6.3	
Wimereux	61.7 \pm 5.8	27.7 \pm 6.2	5.0 \pm 2.3	5.7 \pm 2.2	= 0.6	89.3 \pm 2.4	= 0.05
Sidi Fredj	60.7 \pm 8.1	19.8 \pm 2.6	10.7 \pm 3.1	8.8 \pm 4.2	= 0.6	80.5 \pm 6.8	= 0.4
Tamentfoust	62.2 \pm 6.6	20.2 \pm 3.9	8.2 \pm 3.3	9.5 \pm 2.6	= 0.6	82.3 \pm 4.8	= 0.5
Algiers Beach	20.5 \pm 17.4	47.8 \pm 13.5	24.3 \pm 7.9	7.3 \pm 3.3	< 0.0005	68.3 \pm 6.4	< 0.0005

N: normal plutei; R: retarded plutei; P1: abnormal plutei; P2; Blastula; V: viable plutei; Du1: result of the Dunnett test comparing the rate of normal plutei to the control; Du2: result of the Dunnett test comparing the rate of viable plutei to the control.

Table 6. Comparison of the mean concentrations ($\mu\text{g g}^{-1}$ dw) of metals in the gonads of *Paracentrotus lividus*.

Authors (sampling date)	Sites	Zn	Pb	Cu	Cd	Fe
Present study March 2002	Algiers Beach					
	Females	385.5	6.14	2.84	0.14	73.8
	Males	32.9	7.78	3.19	0.08	19.3
	Tamentfoust					
	Females	538.2	1.5	2.49	0.12	113
	Males	76.1	0.88	3.88	0.05	112.6
	Sidi Fredj					
	Females	366.9	0.68	3.42	0.14	71.1
Males	52.9	0.90	4.42	0.05	92.7	
Bayed et al 2005 March 2000	Atlantic Morocco					
	Rabat					
	Females		<u>35.32</u>	3.34	<u>25.15</u>	
	Males		12.41	0.52	2.58	
	Bouznika					
	Females		11.3	<u>5.56</u>	2.21	
	Males		7.18	1.52	1.51	
	Mohammedia					
Females		5.22	2.51	2.24		
Males		6.14	1.18	2.32		
Warnau et al 1998	Calvi	124	2.25	0.15	3.47	51
	Ischia	140	3.02	0.41	3.41	90
	Marseille	109	3.68	0.19	3.51	39
Storelli et al 2001 April 1998	Adriatic Sea	157.1	0.86	0.24	5.19	18.37

The higher values are in bold and underlined